

# Differential Effects of Source-Specific Particulate Matter on Emergency Hospitalizations for Ischemic Heart Disease in Hong Kong

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Differential Effects of Source-Specific Particulate Matter on **Emergency Hospitalizations for Ischemic Heart Disease in Hong** 

Kong

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Running title: Effects of PM<sub>10</sub> sources on ischemic heart disease risks

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## **Abstract**

**Background**: Ischemic heart disease (IHD) is a major public health concern. Although many epidemiologic studies have reported evidence of adverse effects of particulate matter (PM) mass on IHD, significant knowledge gaps remain regarding the potential impacts of different PM sources. Like PM size, PM sources may influence toxicological characteristics.

**Objectives**: We identified contributing sources to  $PM_{10}$  mass and estimated the acute effects of  $PM_{10}$  sources on daily emergency IHD hospitalizations in Hong Kong.

**Methods**: We analyzed the concentration data of nineteen  $PM_{10}$  chemical components measured between 2001 and 2007 by positive matrix factorization to apportion  $PM_{10}$  mass, and used generalized additive models to estimate associations of inter-quartile range (IQR) increases in  $PM_{10}$  exposures with IHD hospitalization for different lag periods (up to 5 days), adjusted for potential confounders.

**Results**: Eight PM<sub>10</sub> sources were identified: vehicle exhaust, soil/road dust, regional combustion, residual oil, fresh sea salt, aged sea salt, secondary nitrate and secondary sulfate. Secondary sulfate, vehicle exhaust, and secondary nitrate contributed over half of the PM<sub>10</sub> mass. While associations with IQR increases in two-day moving averages (lag<sub>01</sub>) were statistically significant for most sources based on single-source models, only PM<sub>10</sub> from vehicle exhaust (1.87%; 95% CI: 0.66, 3.10; IQR = 4.9  $\mu$ g/m³), secondary nitrate (2.28%; 95% CI: 1.15, 3.42; IQR = 8.6  $\mu$ g/m³) and aged sea salt (1.19%; 95% CI: 0.04, 2.36; IQR = 5.9  $\mu$ g/m³) were significantly associated with IHD hospitalizations in multi-source model. Analysis using chemical components provided similar findings.

Conclusion: Emergency IHD hospitalization was significantly linked with  $PM_{10}$  from vehicle exhaust, nitrate-rich secondary PM, and sea salt-related PM. Findings may help prioritize toxicological research and guide future monitoring and emission control polices.

### Introduction

Over the past decades, epidemiologic evidence has linked ambient particulate matter (PM) pollution to increased cardiovascular morbidity and mortality (Dominici et al. 2006; Peng et al. 2009). Of the cardiovascular endpoints, ischemic heart disease (IHD) is a major public health concern. IHD is defined as a narrowing of the coronary vessels that supply blood to the heart muscle. It was the leading cause of death worldwide in 2008 and the second leading cause of death in Hong Kong, with a population around 7 million and a daily average of 12 IHD deaths in 2011 (Department of Health HKSAR 2013). Evidence from the U.S. and Europe of increases in IHD events after acute exposure to elevated PM concentrations has been convincing (Dominici et al. 2006; Forastiere et al. 2005; Pope et al. 2006). It has been estimated that an average 10 µg/m<sup>3</sup> reduction in PM<sub>2.5</sub> in 204 U.S. counties would prevent over 1,500 IHD hospitalizations per year (Dominici et al. 2006). However, previous studies in Hong Kong did not observe associations between PM and IHD hospitalizations (Wong CM et al. 2002, Wong et al. 1999) or mortality upon adjusting for gaseous pollutants (Wong TW et al. 2002). Heterogeneity in findings may reflect the fact that PM is a complex mixture of particles that vary in physical attributes, chemical composition, solubility and emission sources (Pope and Dockery 2006).

Growing research emphasis has been placed on PM sources and chemical composition (Health Effects Institute 2002; U.S. National Research Council 2004). Since PM sources generate mixtures of air pollutants with different physicochemical compositions, they might affect the relative toxicity of PM. This hypothesis is supported by toxicological evidence suggesting that PM-induced biologic effects depended on the zone of origin (e.g. industrial zone) (Alfaro-Moreno et al. 2002). Currently, the majority of studies have associated IHD, especially

myocardial infarction, with traffic-related pollution exposures estimated using surrogate pollutants (e.g., PM<sub>2.5</sub> mass, carbon monoxide, nitrogen dioxide) or direct exposure data (e.g. time spent in traffic) (D'Ippoliti et al. 2003; Lanki et al. 2006b; Peters et al. 2004). Despite these findings, it has been a challenge to quantitatively assess the impacts of multiple PM emission sources on IHD. Associations with PM sources have been inconsistent across existing studies. While some have reported associations of traffic-related and/or combustion-generated PM with increases in repolarization, inflammatory markers, and ST segment depressions among IHD patients (Lanki et al. 2006a; Yue et al. 2007), others have reported that IHD hospitalizations were not linked with traffic-related particles or other PM sources (Halonen et al. 2009; Lall et al. 2011).

In Hong Kong, although research on PM pollution and health outcomes has been active since the late–1990s, specific PM chemical components and sources responsible for the adverse effects have rarely been investigated. In this study, we took advantage of the PM<sub>10</sub> speciation data that have been available for over a decade to identify contributing sources to PM<sub>10</sub> mass using a source apportionment model, and estimate the acute effects of PM<sub>10</sub> sources on daily emergency IHD hospital admissions.

### Methods

### Data

The Hong Kong Environmental Protection Department has been collecting 24-hour filter samples of PM<sub>10</sub> regularly at six general and one roadside air quality monitoring stations since 2001 (Yuan et al. 2012). These monitoring stations were interspersed in different districts of Hong Kong. We included only data from the six general stations that are not in direct proximity

to traffic, industrial sources, buildings, or residential sources of emissions from the burning of coal, waste or oil. These stations serve to capture the air quality that the general population is exposed to on a regular basis. Twenty-six PM<sub>10</sub> chemical components were speciated from the filter samples via various analytical methods as described in detail elsewhere (Yuan et al. 2012). Speciation data between 1 January 2001 and 31 December 2007 were included in the present study. The PM<sub>10</sub> sampling frequency was on average every-sixth-day, with each station operated on a distinct sampling schedule. On a particular day, there might be none or multiple samples taken across the stations. Overall, 71% of the study days were covered by measurements from at least one station. We obtained daily mean temperature, relative humidity from the Hong Kong Observatory for the same study period.

We acquired daily counts of emergency hospital admissions between 1 January 2001 and 31 December 2007 from the Hong Kong Hospital Authority (Wong et al. 1999). Data were coded according to the 9th revision of the International Classification of Diseases (ICD–9; World Health Organization 1975). Hospitalizations for IHD (ICD–9:410–414) were extracted to construct the time series. Hospitalizations due to influenza (ICD–9:487) were extracted and treated as a potential confounder in the regression analysis.

### Statistical analysis

We first used the U.S. Environmental Protection Agency's positive matrix factorization (PMF) version 3.0 to identify a set of factors, interpreted as emission sources, and to estimate the source contributions to  $PM_{10}$  mass (Hopke 2008). Station-specific measurements of elemental carbon (EC), organic matter (OM), nitrate ( $NO_3^-$ ), sulfate ( $SO_4^{2-}$ ), ammonium ion ( $NH_4^+$ ), chloride ion ( $CI_1^-$ ), sodium ion ( $Na_4^+$ ), potassium ion ( $K_1^+$ ), aluminum (Al), arsenic (As), calcium (Ca),

cadmium (Cd), iron (Fe), magnesium (Mg), manganese (Mn), nickel (Ni), lead (Pb), vanadium (V) and zinc (Zn) of PM<sub>10</sub> were entered into the PMF model. Details on PMF modeling have been described elsewhere (Reff et al. 2007).

We removed the station-specific influence on the resultant concentrations of each PM<sub>10</sub> source by: 1) computing the annual mean concentration ( $X_i$ ) for each monitoring station i; 2) subtracting the annual mean from the daily mean concentration for station i on each sample day j ( $X_{ij}$ ); 3) adding the annual mean of all stations (X) to the resulting centered values ( $X_{ij} - X_i$ ) for each station and sampling day to produce  $X''_{ij} = X_{ij} - X_i + X$ ; and 4) taking the average of  $X''_{ij}$  over all stations (Wong et al. 2001). The final PM<sub>10</sub> sources time-series contained non-missing territory-wide mean concentrations of PM<sub>10</sub> sources for 1,805 days (71% of the 2,556 total days), which is about five days a week. All pollutant concentrations were expressed in  $\mu$ g/m³, except for EC and OM reported in microgram of carbon in cubic meter ( $\mu$ gC/m³).

Generalized additive models with log link and Poisson error were used to estimate the associations between PM<sub>10</sub> sources and emergency IHD hospital admissions (Hastie and Tibshirani 1990). We adopted *a priori* model specification to guide the selection of degrees of freedom (df) for time-varying variables: smoothing splines with 8 df per year for time trend, 6 df for current day temperature and previous 3 days moving average, and 3 df for current day relative humidity and previous 3 days moving average (Bell et al. 2009; Peng et al. 2009). We included dummy variables for day of week, public holidays, and influenza epidemics (Wong CM et al. 2002).

We investigated the possible lag distribution of associations with each  $PM_{10}$  source for exposures on the same day (lag<sub>0</sub>) and for daily exposures on the previous 1 to 5 days (lag<sub>1</sub> to lag<sub>5</sub>).

However, we focused primarily on the two-day moving average of exposure on the same day and the previous day (lag<sub>01</sub>) a priori based on previous studies (Wong et al. 2008; Wong CM et al. 2002). Furthermore, we conducted multi-source analysis to estimate mutually adjusted effects of PM<sub>10</sub> sources on emergency IHD hospitalizations (Ostro et al. 2011). To minimize multicollinearity, we used backward elimination with an exclusion criterion of p > 0.10 to select PM sources to include in the final multi-source model, while controlling for time trend, seasonality, meteorological conditions, calendar effects and influenza epidemic. Pearson's correlations were used to summarize the relationships between source-apportioned PM<sub>10</sub>. PM<sub>10</sub> "tracer" components, which are characterized as the typical components that are exclusively or largely derived from a particular source, were also examined, and those tracers that are found specifically in the sources included in the final multi-source model were further tested in a separate multi-pollutant model to validate the multi-source findings. A smoothing function with 3 df was applied to graphically describe the relationships between sources and IHD hospitalizations, while adjusting for time-varying confounders. For sensitivity analysis, we repeated the time-series analysis after either imputing source concentrations for the days without samples from any stations (751 days) by linear interpolation using the na.approx function in the R zoo package, or replacing the missing data with non-missing measurement values from the previous day. Moreover, we evaluated the impact of alternative df values (5-12) for time trend on the risk estimates. All estimates were reported as the percent increase [(relative risk-1)×100%] in daily emergency IHD hospital admissions for an inter-quartile range (IQR) increment in pollutant concentrations. Where appropriate, 95% confidence intervals were calculated. We performed all time-series analyses in the statistical environment R Software, version 2.15.0 (R Development Core Team, Vienna).

### Results

Eight PM<sub>10</sub> sources were identified, namely vehicle exhaust, soil/road dust (e.g., from exposed soil, unpaved roads), regional combustion, residual oil combustion (fuel emissions from marine vessels), fresh sea salt, aged sea salt, secondary nitrate and secondary sulfate. Figure 1 shows the estimated PM<sub>10</sub> source profiles, depicted as explained variations (EV) that indicate the relative contribution of each source to the variation of a given chemical component (Paatero and Tapper 1994). For instance, vehicle exhaust emission accounted for 80% of the variation in EC. Regional combustion emission was identified as a composite of two sources, which were wood/biomass burning based upon the abundance (i.e., large EV) of K<sup>+</sup>, and coal combustion in power plants and industrial facilities in the adjacent Pearl River Delta region based upon the abundances of As, Cd, Pb and Zn in the source profile, that cannot be further separated (Yuan et al. 2012). Table 1 summarizes the levels of PM<sub>10</sub> pollution, weather conditions, and IHD hospital admission counts. Between 2001 and 2007, the daily average concentration for PM<sub>10</sub> in Hong Kong was  $55.8 \pm 32.5 \text{ µg/m}^3$ . Secondary sulfate accounted for the largest fraction of total PM<sub>10</sub> mass (23.6%), followed by vehicle exhaust (15.1%) and secondary nitrate (14.9%). The mean daily average temperature and relative humidity were 23.6°C and 78.3% respectively (Table 1). During the study period, there were 76,659 hospitalizations for IHD ( $30 \pm 7$  admissions per day). Single-source models of single-day exposure lags showed similar patterns of associations for most of the PM<sub>10</sub> sources, in that IHD hospitalizations were positively associated with exposure on the same day (lag<sub>0</sub>), maximal for lag<sub>0</sub> or lag<sub>1</sub>, and lowest at later lags (lag<sub>4</sub>-lag<sub>5</sub>) (Figure 2). At lag<sub>01</sub> (Figure 3A), the source that was most strongly associated with IHD hospitalizations was secondary nitrate (2.89% increase; 95% CI: 1.83, 3.95; IQR = 8.6 µg/m<sup>3</sup>), followed by vehicle exhaust (2.35%; 95% CI: 1.24, 3.47; IQR = 4.9  $\mu$ g/m³) and regional combustion (2.26%; 95% CI: 0.98, 3.55; IQR = 11.7  $\mu$ g/m³), after adjusting for time trend, seasonality, meteorological conditions, calendar effect and influenza epidemic. Significant positive associations were also found for particles originated from soil/road dust (per 6.9  $\mu$ g/m³), residual oil (per 2.2  $\mu$ g/m³) and secondary sulfate (per 15.8  $\mu$ g/m³), corresponding to estimated increases in IHD hospitalizations of 0.97%–1.37%.

Correlations between source-apportioned PM<sub>10</sub> were nil-to-moderate. The highest correlation coefficient was 0.67 between regional combustion and secondary sulfate, followed by that between regional combustion and secondary nitrate (0.59; Table 2). Backward elimination resulted in a multi-source regression model that included vehicle exhaust, aged sea salt, and secondary nitrate sources only (Figure 3B). All other sources, though statistically significant in single-source models, were eliminated from the final multi-source model based on p > 0.10. The estimated effects of vehicle exhaust (1.87%; 95% CI: 0.66, 3.10) and secondary nitrate (2.28%; 95% CI: 1.15, 3.42) at lag<sub>01</sub> were slightly attenuated relative to single-source model estimates, but remained significant in the final multi-source model. The association between aged sea salt and IHD hospitalizations was stronger based on the multi-source model (1.19%; 95% CI: 0.04, 2.36;  $IQR = 5.9 \mu g/m^3$ ) compared with the single-source model (0.68%; 95% CI: -0.43, 1.79). Associations with "tracer" components of secondary nitrate and vehicle exhaust were similar, that is, NO<sub>3</sub> (tracer for secondary nitrate) and EC (tracer for vehicle exhaust) were associated with 1.95% (95% CI: 0.68, 3.25;  $IQR = 3.4 \mu g/m^3$ ) and 1.67% (95% CI: 0.58, 2.78; IQR = 1.6μg/m<sup>3</sup>) increase in IHD hospitalizations, respectively (Figure 3C).

We examined the concentration-response (C-R) relations for vehicle exhaust, aged sea salt and secondary nitrate in a multi-source model. We observed moderate positive relationships over the IQRs of source concentrations, except for aged sea salt where a neutral relationship was seen (data not shown). The risk estimates were not sensitive to alternative time-series models in which we imputed missing data (data not shown). Varying df for time trend (5-12 per year, data not shown) did not substantially change the regression results either.

### **Discussion**

Research directly delineating the health impacts of PM emission sources is relatively limited. Most studies rely on ambient concentrations of a PM chemical component as surrogate of the combined exposure to one source (Stanek et al. 2011). This a priori selection can be complicated when interpreting the results, because many components are emitted from numerous sources, and the same component may not serve as tracer to the same source at different locations (Sarnat et al. 2008). We joined a small but growing number of epidemiologic studies to conduct source apportionment analysis and quantitatively estimate the associations between multiple PM sources and health outcome. Early short-term air pollution studies conducted in western countries identified some associations between PM sources and mortality (Cakmak et al. 2009b; Halonen et al. 2009; Ito et al. 2006; Laden et al. 2000; Mar et al. 2000, 2006; Ostro et al. 2011), and gradually, researchers have also linked certain PM sources to hospital admissions (Andersen et al. 2007; Cakmak et al. 2009a; Halonen et al. 2009; Lall et al. 2011; Sarnat et al. 2008). Overall, these studies have reported some evidence suggesting that PM sources representing traffic/motor exhausts, regional/secondary sulfate, and coal/oil combustion may be more toxic than other PM sources, as summarized in a recent review by Stanek et al. (2011). Nonetheless, there is

insufficient evidence to draw more specific conclusions across studies. Since emission sources of air pollutants vary not only temporally, but also geographically, studies on source-apportioned PM mass under different atmospheres are needed to improve our understanding of PM-related health effects.

To our knowledge, this is the first Asian study to investigate the health impacts of multiple PM sources. We estimated the associations of short-term exposure to source-apportioned PM<sub>10</sub> mass with emergency IHD hospital admissions in Hong Kong, a coastal urban city on the boundary region of Asian continent and Pacific Ocean. In contrast to previous studies conducted in New York City and Helsinki, Finland that reported no associations of PM<sub>2.5</sub> sources with IHD hospital admissions (Halonen et al. 2009; Lall et al. 2011), we observed significant associations between IQR increases in several PM<sub>10</sub> sources and daily IHD hospitalizations for single-day lag periods up to 5 days prior. Differences between our findings and those of previous studies might be related to the longer study period (7 years), larger combined sample size in the present study, as well as the differences in pollution compositions and population susceptibility between cities. Although precise patho-physiological mechanisms connecting ambient air pollution with IHD remain to be determined, it is commonly hypothesized that PM sources may trigger and/or enhance the formation of reactive oxygen species that induce inflammation, the formation of atherosclerotic plaques as well as vaso-constriction, resulting in reduced oxygen supply of heart tissues, and thereby leading to IHD (Lawal and Araujo 2012; Peters 2011).

Secondary nitrate in  $PM_{10}$  (per 8.6  $\mu g/m^3$ ) was associated with the largest increases in IHD hospitalizations at  $lag_{01}$  in both single-source and multi-source models. On the other hand, the estimated association of secondary sulfate diminished after adjusting for other sources, and

secondary sulfate was dropped from the final multi-source model based on the backward elimination criterion (data not shown). This is somewhat surprising considering that secondary sulfate accounted for the largest fraction of total PM<sub>10</sub> in Hong Kong. While most studies that examined these associations found that sulfate-rich secondary PM was more strongly associated with mortality and hospital admissions than nitrate (e.g., Ito et al. 2006; Mar et al. 2000, 2006; Sarnat et al. 2008), a few studies have reported that nitrate in PM<sub>2.5</sub>, rather than sulfate, was significant predictor of mortality (Fairley 1999; Ostro et al. 2011). Secondary nitrate and secondary sulfate derive largely from the oxidation of nitrogen oxides and sulfur dioxide emitted from combustion of fossil fuels respectively. Whilst both are acidic in nature, their strength of acidity varies greatly depending upon the city-specific interactions between local emissions, regional transports, and meteorological conditions (Schlesinger and Cassee 2003). Although animal toxicological evidence is inconclusive, it is hypothesized that acidic aerosols may lower the pH within the airways by depositing hydrogen ions, thereby triggering adverse reactions (Kelly and Fussell 2012). In China, strong economic growth and high total energy consumption have led to substantial increases in anthropogenic nitrogen and sulfur emissions over the past decades (Liu et al. 2013; Lu et al. 2010). Studies showed that emissions of nitrogen oxides and sulfur dioxide in the adjacent Pearl River Delta region due to rapid industrialization and urbanization have been the dominant contributors to secondary nitrate and secondary sulfate in Hong Kong through regional transportation (Guo et al. 2009; Yuan et al. 2012). Our finding on secondary nitrate is of particular importance for lending urgency to policy-makers, particularly in developing economies, regarding both local and regional emission control and reduction of gaseous pollutants.

We found that an IQR increment (4.9  $\mu g/m^3$ ) in two-day moving average concentration of PM<sub>10</sub> from vehicle exhaust was associated with a 1.91% estimated increase (95% CI: 0.70, 3.13%) in IHD hospital admissions after adjusting for other statistically significant sources. In accordance with these results, EC as a chemical tracer of vehicle exhaust (largely from diesel engine) was also significantly associated with IHD hospitalization risk. Vehicle exhaust-related PM are combustion-derived particles that primarily accumulate in the fine fraction of PM<sub>10</sub>. Previous epidemiologic studies have reported that mobile sources PM<sub>2.5</sub> and EC are stronger predictors of overall cardiovascular outcomes than other sources and components (Cakmak et al. 2009a, 2009b; Lall et al. 2011; Mar et al. 2006; Ostro et al. 2011; Sarnat et al. 2008), which is consistent with our finding of an association of IHD with vehicle exhaust PM. Plausible biological mechanisms include elevated levels of inflammatory biomarkers, impaired endotheliumdependent vasodilation, and promotion of ST-segment depression (Dales et al. 2007; Lanki et al. 2006a; Yue et al. 2007). In Hong Kong, where road density was among the highest in the world at 254 vehicles per kilometer of road in 2009, exposure to traffic-related air pollution is ubiquitous (World Bank 2012). These findings on vehicle exhaust particles stress the importance of continuous reduction of overall traffic and related emissions, and configuration of urban environment to reduce personal exposure to traffic.

We observed that aged sea salt was associated with an increased risk of IHD hospitalizations after adjusting for other sources. Sea salts are most abundantly found in the coarser fraction of PM<sub>10</sub>. While Mar et al. (2006) reported that sea salt was consistently associated with elevated cardiovascular and total mortality in Atlanta, Georgia across the various source apportionment analyses in an inter-method comparison study, most epidemiologic studies that estimated this

association found no relationship between sea salt and health outcomes (Andersen et al. 2007; Gent et al. 2009; Lanki et al. 2006a; Ostro et al. 2011). Several studies suggested possible relations between sea salt and secondary nitrate, as nitric acid may react with marine particles to form coarse mode nitrate along coastal areas (Lee et al. 2008; Zhuang et al. 1999). However, we observed null-to-weak correlations between secondary nitrate and sea salts. Such association should be investigated further.

Our findings add to the existing literature in several ways. First, we examined air pollution association with specific cardiovascular endpoint, as opposed to a broad composite endpoint of different cardiovascular events, to provide better insight into the plausible biologic mechanisms. Second, with nearly 80,000 IHD hospital admissions over 7 years, our study was well powered to detect statistically significant associations. Moreover, this was one of the few epidemiologic studies that focused on exposure to source-apportioned PM<sub>10</sub>, while most existing studies were based on source-apportioned PM<sub>2.5</sub>. This allowed us to identify adverse associations of not only sources that primarily generate finer mode PM<sub>10</sub>, but also those that produce coarser mode PM<sub>10</sub>.

Although we provided evidence of the health impacts of several PM<sub>10</sub> sources in Hong Kong, these findings should be interpreted with caution. While backward elimination procedure was used to identify a subset of predictors with the most statistically significant relationship with IHD hospitalization, this approach might not guarantee a truly "best" reduced model (Breiman 1996). The importance of PM<sub>10</sub> sources (e.g., regional combustion, secondary sulfate) excluded from the final multi-source model should not be diminished, as the statistical elimination procedure does not indicate or account for biological importance. Another limitation of this study was the every-sixth-day sampling scheme for the PM<sub>10</sub> speciation data, resulting in nearly one-third of

study days without samples from any stations. Exposure misclassification error might exist, however, our risk estimates were insensitive to alternative interpolation methods (data not shown). Moreover, PM from local emissions (e.g., vehicle exhaust, soil/road dusts) tends to have more error than regional PM sources (e.g., secondary PM), given their higher spatial heterogeneity (Ito et al. 2004). Such issues of representativeness associated with PM sources may hinder the interpretations of the relative strengths of the observed associations in monitor-based studies of ambient PM pollution.

### Conclusion

We report evidence that PM<sub>10</sub> from vehicle exhaust, nitrate-rich secondary PM, and sea salt-related PM were significantly associated with elevated IHD hospitalization risks in Hong Kong. This study joins a growing body of literature to report evidence of adverse effects of source-apportioned PM mass, which would help prioritize research on the biologic mechanisms linking PM pollution to cardiac events and guide future monitoring and emission control polices.

### References

- Alfaro-Moreno E, Martínez L, García-Cuellar C, Bonner JC, Murray JC, Rosas I, et al. 2002. Biologic effects induced in vitro by PM10 from three different zones of Mexico City. Env. Heal. Perspect 110:715–720.
- Andersen ZJ, Wahlin P, Raaschou-Nielsen O, Scheike T, Loft S. 2007. Ambient particle source apportionment and daily hospital admissions among children and elderly in Copenhagen. J Expo Sci Env. Epidemiol 17: 625–636.
- Bell ML, Ebisu K, Peng RD, Samet JM, Dominici F. 2009. Hospital admissions and chemical composition of fine particle air pollution. Am J Respir Crit Care Med 179: 1115–1120.
- Breiman L. 1996. Heuristics of instability and stabilization in model selection. Ann Stat 24:2350–2383.
- Cakmak S, Dales R, Gultekin T. 2009a. Components of particulate air pollution and emergency department visits in Chile. Arch Env. Occup Heal. 64:37–41.
- Cakmak S, Dales RE, Vida CB. 2009b. Components of particulate air pollution and mortality in Chile. Int J Occup Env. Heal. 15:152–158.
- D'Ippoliti D, Forastiere F, Ancona C, Agabiti N, Fusco D, Michelozzi P, et al. 2003. Air pollution and myocardial infarction in Rome: a case-crossover analysis. Epidemiology 14: 528–535.
- Dales R, Liu L, Szyszkowicz M, Dalipaj M, Willey J, Kulka R, et al. 2007. Particulate air pollution and vascular reactivity: the bus stop study. Int Arch Occup Env. Heal. 81: 159–164.
- Department of Health HKSAR. 2013. HealthyHK: Major Causes of Death. Available: http://www.healthyhk.gov.hk/phisweb/en/healthy\_facts/disease\_burden/major\_causes\_death/major\_causes\_death/ [accessed 25 May 2013].
- Dominici F, Peng RD, Bell ML, Pham L, McDermott A, Zeger SL, et al. 2006. Fine particulate air pollution and hospital admission for cardiovascular and respiratory diseases. JAMA 295: 1127–1134.
- Fairley D. 1999. Daily mortality and air pollution in Santa Clara County, California: 1989-1996. Env. Heal. Perspect 107:637–641.

- Forastiere F, Stafoggia M, Picciotto S, Bellander T, D'Ippoliti D, Lanki T, et al. 2005. A case-crossover analysis of out-of-hospital coronary deaths and air pollution in Rome, Italy. Am J Respir Crit Care Med 172: 1549–1555.
- Gent JF, Koutrakis P, Belanger K, Triche E, Holford TR, Bracken MB, et al. 2009. Symptoms and medication use in children with asthma and traffic-related sources of fine particle pollution. Env. Heal. Perspect 117: 1168–1174.
- Guo H, Ding AJ, So KL, Ayoko G, Li YS, Hung WT. 2009. Receptor modeling of source apportionment of Hong Kong aerosols and the implication of urban and regional contribution. Atmos Env. 43:1159–1169.
- Halonen JI, Lanki T, Yli-Tuomi T, Tiittanen P, Kulmala M, Pekkanen J. 2009. Particulate air pollution and acute cardiorespiratory hospital admissions and mortality among the elderly. Epidemiology 20: 143–153.
- Hastie TJ, Tibshirani RJ. 1990. Generalized Additive Models. Chapman & Hall/CRC.
- Health Effects Institute. 2002. Understanding the health effects of components of the particulate matter mix: progress and next steps. Available: http://pubs.healtheffects.org/getfile.php?u=244 [accessed 26 May 2013].
- Hopke PK. 2008. The use of source apportionment for air quality management and health assessments. J Toxicol Env. Heal. A 71: 555–563.
- Ito K, Christensen WF, Eatough DJ, Henry RC, Kim E, Laden F, et al. 2006. PM source apportionment and health effects: 2. An investigation of intermethod variability in associations between source-apportioned fine particle mass and daily mortality in Washington, DC. J. Expo. Sci. Environ. Epidemiol. 16: 300–310.
- Ito K, Xue N, Thurston G. 2004. Spatial variation of PM2.5 chemical species and source-apportioned mass concentrations in New York City. Atmos Env. 38: 5269–5282.
- Kelly FJ, Fussell JC. 2012. Size, source and chemical composition as determinants of toxicity attributable to ambient particulate matter. Atmos Env. 60: 504–526.
- Laden F, Neas LM, Dockery DW, Schwartz J. 2000. Association of fine particulate matter from different sources with daily mortality in six U.S. cities. Env. Heal. Perspect 108:941–947.
- Lall R, Ito K, Thurston GD. 2011. Distributed lag analyses of daily hospital admissions and source-apportioned fine particle air pollution. Env. Heal. Perspect 119: 455–460.

- Lanki T, Hartog JJ de, Heinrich J, Hoek G, Janssen N a. H, Peters A, et al. 2006a. Can we identify sources of fine particles responsible for exercise-induced ischemia on days with elevated air pollution? The ULTRA study. Env. Heal. Perspect 114: 655–660.
- Lanki T, Pekkanen J, Aalto P, Elosua R, Berglind N, D'Ippoliti D, et al. 2006b. Associations of traffic related air pollutants with hospitalisation for first acute myocardial infarction: the HEAPSS study. Occup Env. Med 63: 844–851.
- Lawal AO, Araujo JA. 2012. Particulate matter and cardiovascular health effects. In *Air pollution Monitoring, modelling and health* (Dr. Mukesh Khareed.), p. 386, InTech.
- Lee T, Yu X, Ayres B, Kreidenweis S, Malm W, Collettjr J. 2008. Observations of fine and coarse particle nitrate at several rural locations in the United States. Atmos Env. 42: 2720–2732.
- Liu X, Zhang Y, Han W, Tang A, Shen J, Cui Z, et al. 2013. Enhanced nitrogen deposition over China. Nature 494: 459–462.
- Lu Z, Streets DG, Zhang Q, Wang S, Carmichael GR, Cheng YF, et al. 2010. Sulfur dioxide emissions in China and sulfur trends in East Asia since 2000. Atmos Chem Phys 10: 6311–6331.
- Mar TF, Ito K, Koenig JQ, Larson T V, Eatough DJ, Henry RC, et al. 2006. PM source apportionment and health effects. 3. Investigation of inter-method variations in associations between estimated source contributions of PM2.5 and daily mortality in Phoenix, AZ. J Expo Sci Env. Epidemiol 16: 311–320.
- Mar TF, Norris G a, Koenig JQ, Larson T V. 2000. Associations between air pollution and mortality in Phoenix, 1995-1997. Env. Heal. Perspect 108:347–353.
- Ostro B, Tobias A, Querol X, Alastuey A, Amato F, Pey J, et al. 2011. The Effects of Particulate Matter Sources on Daily Mortality: a case-crossover study of Barcelona, Spain. Env. Heal. Perspect 119:1781–1787.
- Paatero P, Tapper U. 1994. Positive matrix factorization: A non-negative factor model with optimal utilization of error estimates of data values. Environmetrics 5: 111–126.
- Peng RD, Bell ML, Geyh AS, McDermott A, Zeger SL, Samet JM, et al. 2009. Emergency admissions for cardiovascular and respiratory diseases and the chemical composition of fine particle air pollution. Env. Heal. Perspect 117: 957–963.

- Peters A. 2011. Air pollution and ischemic heart disease. In *Environmental Cardiology: Pollution and Heart Disease*, p. 390, Royal Society Of Chemistry, Cambridge, UK.
- Peters A, Klot S Von, Heier M, Trentinaglia I, Hörmann A, Wichmann E, et al. 2004. Exposure to traffic and the onset of myocardial infarction. N Engl J Med 351:1721–1730.
- Pope CA, Muhlestein JB, May HT, Renlund DG, Anderson JL, Horne BD. 2006. Ischemic heart disease events triggered by short-term exposure to fine particulate air pollution. Circulation 114: 2443–2448.
- Pope CAI, Dockery DW. 2006. Health effects of fine particulate air pollution: lines that connect. J Air Waste Manag Assoc 56:709–742.
- Reff A, Eberly SI, Bhave P V. 2007. Receptor modeling of ambient particulate matter data using positive matrix factorization: review of existing methods. J Air Waste Manag Assoc 57:146–154.
- Sarnat JA, Marmur A, Klein M, Kim E, Russell AG, Sarnat SE, et al. 2008. Fine particle sources and cardiorespiratory morbidity: an application of chemical mass balance and factor analytical source-apportionment methods. Env. Heal. Perspect 116:459–466.
- Schlesinger RB, Cassee F. 2003. Atmospheric secondary inorganic particulate matter: the toxicological perspective as a basis for health effects risk assessment. Inhal Toxicol 15: 197–235.
- Stanek LW, Sacks JD, Dutton SJ, Dubois J-JB. 2011. Attributing health effects to apportioned components and sources of particulate matter: An evaluation of collective results. Atmos. Environ. 45: 5655–5663.
- The U.S. National Research Council. 2004. *Research Priorities for Airborne Particulate Matter:*IV. Continuing Research Progress. National Academy Press, Washington (DC).
- Wong CM, Atkinson RW, Anderson HR, Hedley AJ, Ma S, Chau PY-K, et al. 2002. A tale of two cities: effects of air pollution on hospital admissions in Hong Kong and London compared. Env. Heal. Perspect 110:67–77.
- Wong CM, Ma S, Hedley AJ, Lam TH. 2001. Effect of air pollution on daily mortality in Hong Kong. Env. Heal. Perspect 109:335–340.

- Wong CM, Vichit-Vadakan N, Kan H, Qian Z. 2008. Public Health and Air Pollution in Asia (PAPA): a multicity study of short-term effects of air pollution on mortality. Env. Heal. Perspect 116: 1195–1202.
- Wong TW, Lau TS, Yu TS, Neller A, Wong SL, Tam W, et al. 1999. Air pollution and hospital admissions for respiratory and cardiovascular diseases in Hong Kong. Occup Env. Med 56:679–683.
- Wong TW, Tam WS, Yu TS, Wong a HS, Article O. 2002. Associations between daily mortalities from respiratory and cardiovascular diseases and air pollution in Hong Kong, China. Occup Env. Med 59:30–35.
- World Bank. 2012. Vehicles (per km of road). Available: http://data.worldbank.org/indicator/IS.VEH.ROAD.K1/countries [accessed 1 June 2013].
- World Health Organization. 1975. International Classification of Diseases, 9th Revision. Geneva, Switzerland.
- Yuan ZB, Yadav V, Turner JR, Louie PKK, Lau AKH. 2012. Long-term trends of ambient particulate matter emission source contributions and the accountability of control strategies in Hong Kong over 1998–2008. Atmos Env. 1–11.
- Yue W, Schneider A, Stölzel M, Rückerl R, Cyrys J, Pan X, et al. 2007. Ambient source-specific particles are associated with prolonged repolarization and increased levels of inflammation in male coronary artery disease patients. Mutat Res 621: 50–60.
- Zhuang H, Chan CK, Fang M, Wexler AS. 1999. Formation of nitrate and non-sea-salt sulfate on coarse particles. Atmos Env. 33: 4223–4233.

**Table 1.** Descriptive statistics for  $PM_{10}$  sources, meteorological factors and number of emergency hospital admissions in Hong Kong, 2001-2007.

| Variables   | No. of days | Daily mean ± SD | Proportion (%) | IQR  |
|---|-------------|-----------------|----------------|------|
| Emergency hospital admissions (counts)              |             |                 |                |      |
| Ischemic heart diseases (IHD)                       | 2556        | 30 ± 7          |                | 9    |
| Meteorological conditions                           |             |                 |                |      |
| Temperature (°C)                                    | 2556        | $23.6 \pm 4.9$  |                | 8.1  |
| Relative humidity (%)                               | 2556        | $78.3 \pm 9.9$  |                | 11.4 |
| PM <sub>10</sub> concentration (μg/m <sup>3</sup> ) |             |                 |                |      |
| Total PM <sub>10</sub>                              | 1805        | $55.8 \pm 32.5$ | 100            | 44.8 |
| Vehicle exhaust                                     | 1805        | $8.4 \pm 3.7$   | 15.1           | 4.9  |
| Soil/road dust                                      | 1805        | $7.5 \pm 9.0$   | 13.4           | 6.9  |
| Regional combustion                                 | 1805        | $7.5 \pm 9.3$   | 13.5           | 11.7 |
| Residual oil  | 1805        | $2.4 \pm 2.5$   | 4.3            | 2.2  |
| Fresh sea salt                                      | 1805        | $2.1 \pm 2.7$   | 3.7            | 2.0  |
| Aged sea salt                                       | 1805        | $7.2 \pm 4.4$   | 12.8           | 5.9  |
| Secondary nitrate                                   | 1805        | $8.3 \pm 8.8$   | 14.9           | 8.6  |
| Secondary sulfate                                   | 1805        | $13.2 \pm 12.7$ | 23.6           | 15.8 |

Abbreviations: SD, standard deviation; IQR, inter-quartile range.

Table 2. Pearson's correlation among the estimated sources of  $PM_{10}$ .

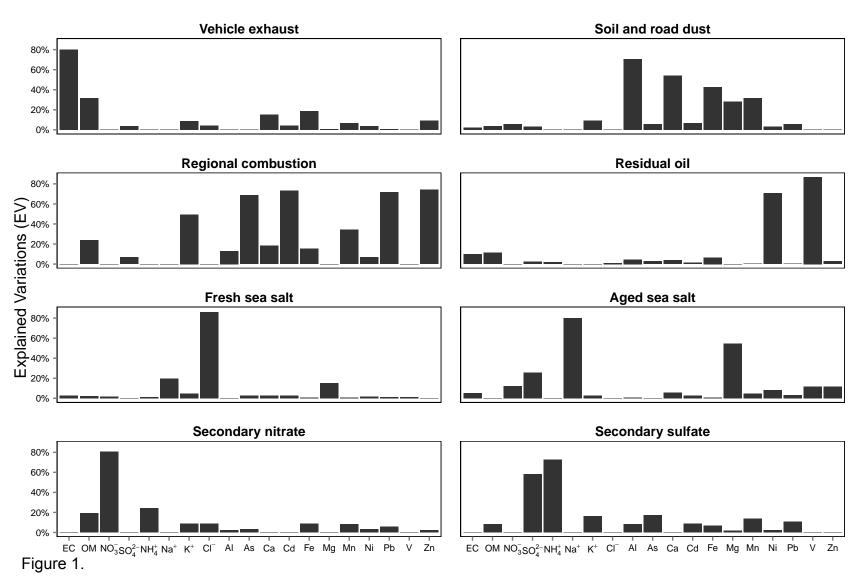
|                     | PM <sub>10</sub> | Vehicle<br>exhaust | Soil/road<br>dust | Regional combustion | Residual<br>oil | Fresh sea<br>salt | Aged sea<br>salt | Secondary nitrate |
|---------------------|------------------|--------------------|-------------------|---------------------|-----------------|-------------------|------------------|-------------------|
| PM <sub>10</sub>    | 1                |                    |                   |                     |                 |                   |                  |                   |
| Vehicle exhaust     | 0.48             | 1                  |                   |                     |                 |                   |                  |                   |
| Soil/road dust      | 0.58             | 0.21               | 1                 |                     |                 |                   |                  |                   |
| Regional combustion | 0.84             | 0.49               | 0.38              | 1                   |                 |                   |                  |                   |
| Residual oil        | 0.40             | 0.35               | -0.02             | 0.29                | 1               |                   |                  |                   |
| Fresh sea salt      | 0.10             | -0.18              | 0.12              | -0.07               | -0.11           | 1                 |                  |                   |
| Aged sea salt       | 0.07             | -0.27              | 0.08              | -0.22               | -0.11           | 0.23              | 1                |                   |
| Secondary nitrate   | 0.76             | 0.34               | 0.29              | 0.59                | 0.34            | 0.24              | 0.05             | 1                 |
| Secondary sulfate   | 0.78             | 0.30               | 0.20              | 0.67                | 0.38            | -0.15             | -0.02            | 0.43              |

# **Figure Legends**

**Figure 1.** PM<sub>10</sub> source profiles, indicated by explained variations (EV) that estimate how much a source explains the variation of a particular chemical component.

**Figure 2.** Percent change (95% CIs) in emergency IHD hospital admissions per IQR increment in PM<sub>10</sub> mass and sources at different lag periods, adjusted for meteorological factors, seasonal and temporal trend, day-of-week and influenza epidemic. See Table 1 for individual IQR values.

**Figure 3.** Percent change (95% CIs) in emergency IHD hospital admissions per IQR increment in two-day moving average concentration (lag<sub>01</sub>) of PM<sub>10</sub> sources based on single-source models (A), multi-source model (B), and PM<sub>10</sub> trace elements based on multi-pollutant model (C). All models were adjusted for time trend, seasonality, meteorological conditions, calendar effects and influenza epidemic. See Table 1 for individual IQR values for sources, and the IQR for EC (tracer for vehicle exhaust), Na<sup>+</sup> (for aged sea salt), and NO<sub>3</sub>- (for secondary nitrate) was 1.6, 1.2 and 3.4 μg/m<sup>3</sup> respectively.



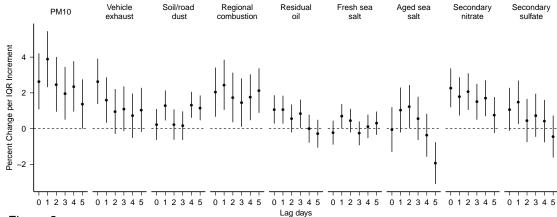


Figure 2

